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Variability in riparian zone potential and actual evapotranspiration in a 1st order agricultural catchment in Southern Ontario, Canada

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Abstract

Micrometeorological and hydrological measurements were made over one growing season using automatic weather stations and weighing lysimeters at several locations within a multiple land-use agricultural catchment in Southern Ontario. This paper compares modelled potential evapotranspiration (PET) and measured actual evapotranspiration (AET) values obtained from the soil weighing lysimeters, and determines the spatial variability in riparian zone AET in a multiple land-use agricultural watershed in Southern Ontario. Two sites were chosen in two different riparian areas of the watershed, representing the surface conditions dominant in the upper and lower reaches of the basin.

The results indicated that AET was higher in the northern end of the basin than in the southern portion of the basin, while the hydrological and energy balance components were similar at both sites. The causes of the different rates are attributed to the surrounding vegetation on adjacent fields and the differing wind regimes.

1. Introduction

Accurate estimates of evapotranspiration (ET) are required, and often lacking, for many management and engineering, field and modelling applications. Further, there is little information about actual ET (AET) from river basins with variable topographies and land use types, typical of temperate agricultural basins (Kotoda, 1989). This lack of information is often due to logistical costs of obtaining accurate AET measurements in large, spatially variable catchments.

Evapotranspiration (ET) is extremely important in the water balance of agricultural watersheds as it can affect stream water export and moisture availability for crops. Knowledge of AET from the field is critical to estimating how much water should be applied to a particular field. However, determining the water quality of streams common in most agricultural basins is the hydrological function of the riparian zones. The

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existence of the stream and associated vegetation depends on groundwater and surface water being accessible to the riparian vegetation. However, there remains little information available on the evaporative loss of these ecologically important strips of riparian vegetation. (Unland et al., 1998).

The riparian zone has been defined as an area where groundwater interacts with surface water and vegetation (Odum, 1971). There are direct and indirect influence of riparian zone ET on a basin's surface and groundwater resources that have been well documented (Goodrich et al., 2000), such as a significant impact on streamflow (Bowie and Kam, 1968; Federer, 1973). Further, quantification of the atmospheric exchange of water through ET from the riparian zone is essential to the basin-wide water balance and the groundwater recharge/depletion process, which make it critical to up-slope moisture status available to crops (Maddock et al., 1998). Although the ET of the riparian zone vegetation is a critical component of the overall catchment water balance, it is still poorly understood (Chehbouni et al., 2000). Much of this lack of information is due none of the typical micrometeorological approaches for measuring ET being appropriate for the narrow riparian corridor (Chehbouni et al., 2000). One approach to overcoming this challenge is to base AET measurements on some estimation of evaporative power or demand (potential or potential equilibrium evapotranspiration) (Hillel, 1998).

As such, many studies are concerned with potential evapotranspiration (PET), which is the maximum rate of ET from a vegetated catchment under the condition of unlimited moisture supply and without advection or heat storage effects (Thomas, 2000; Jacobs et al., 2002). Estimating evaporation by this approach has the advantages of being relatively simple and direct, and makes use of easily measured atmospheric, soil and vegetation variables (Granger and Gray, 1989). Furthermore, PET is a commonly used approach to provide efficient, cost effective estimates of evaporative losses for agricultural and engineering applications (Brutsaert, 1988).

There are a variety of factors that influence the rates of AET and PET. Temperature is significant in that it determines the vapour holding capacity of a parcel of air. A related

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factor is humidity, which can be a significant control on the contribution of plant transpiration to AET through its influences on stomatal activity (Hornberger et al., 1998). If the overlying atmosphere is close to saturated then stomata will close limiting moisture loss from the plants (Jones, 1994).

Wind also has an influence on both ET and PET (Hanson, 1991). Wind flow over a surface draws water molecules away from that surface and moves the water molecules with greater efficiency than calm air. For example, an 8 km h^{-1} wind can increase still-air evapotranspiration by as much as 20% and a 24 km h^{-1} wind by 50% (Hanson, 1991). Finally, the albedo of the surface may also be a significant influence through its control on net radiation, and therefore available energy. Therefore, the rate of evaporation is controlled by: (1) the availability of both water and energy; (2) the strength of the surface-to-air vapour gradient; and (3) the intensity of turbulent motion (Brutsaert, 1984).

The physical characteristics of different land use types will influence these controls and as such rates of ET. However, plants with similar physical characteristics and approximately the same spectral reflectance exhibit similar PET rates, irrespective of plant or soil type (Thornthwaite and Hare, 1965). In most agricultural catchments, in addition to a range in vegetated surfaces, there can exist a significant amount of bare ground, which can account for as much as 50% of the area (Seyfried, 2001). On such terrain, ET will be limited by soil water availability as well as the distribution characteristics of plant species at any given site (Seyfried, 2001). Thus, in a system dominated by agricultural activities, ET values from different terrain elements that comprise that system can differ greatly. This is especially the case when comparing evaporative losses from riparian areas relative to those that are cropped. Quantifying riparian zone ET is essential to developing a better understanding of stream-aquifer-vegetation interactions, improving our ability to simulate these systems (MacNish et al., 2000). That is, accurate estimates of ET in the riparian zone is essential to quantifying biogeochemical processes within this zone.

This paper uses the Priestley-Taylor method for estimating PET and AET values ob-

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tained from soil weighing lysimeters to determine coefficients representative of riparian zone surfaces, and determines the spatial variability of AET within the apparently homogenous riparian zone of an agricultural watershed.

2. Study site

5 This study was conducted in the Strawberry Creek watershed, a small multiple land-use agricultural watershed located in southern Ontario. Strawberry Creek is a perennial, first order stream located in Maryhill, Ontario, approximately 20 km north east of Waterloo, Ontario (42°33' N, 80°23' W) (Fig. 1). The creek is approximately 2 km in length and drains a watershed of about 3 km² (Harris, 1999). Strawberry Creek then
10 flows eastward into Hopewell Creek, which subsequently drains into the Grand River through to Lake Erie.

Strawberry Creek is located in a humid continental climate region (Brown, et al., 1980). There are approximately 3200 growing degree-days and 130 frost-free days with a mean annual temperature of 6.7°C (Brown, et al., 1980). The mean annual
15 precipitation is 858 mm, with approximately 16% of precipitation falling as snow. The mean annual PET for this region is approximately 590 mm, and mean annual AET is approximately 558 mm (Brown, et al., 1980).

The Strawberry Creek watershed contains a variety of vegetation types. The agricultural fields contain soybean (*Heterodera glycines*), and corn (*Zea mays*). In the lower
20 portion of the watershed approximately 60% of the arable land has been in fallow for nine years, the remaining fraction is under corn and strawberry cultivation. In the upper portion of the watershed corn and soybean are planted. The source of Strawberry Creek is a deciduous swamp and there are two smaller woodlots within the watershed. The riparian areas that line Strawberry Creek are 10–15 wide and of low gradient.
25 These riparian zones are dominated by tall grasses (*Graminae spp.*) but also contain Canada Goldenrod (*Solidago canadensis*), Yarrow (*Achillea millefolium*), Thistles (*Cirsium spp.*), Common Milkweed (*Asclepias syriaca*), Queen Anne's Lace (*Daucus*

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carota), Common Burdock (*Arctum minus*), and a few trees (primarily *Salix spp.*).

Two measurement sites were established in the watershed (Fig. 1). The first is located in the northern end of the basin, and will be herein referred to as the “Upper site”. Here the vegetation consists of riparian flora and the soils at this site are characterized as a sandy loam with the top 30 cm of soil having an average bulk density of $0.85 \pm 0.16 \text{ g/cm}^3$, with a bulk density of $1.14 \pm 0.16 \text{ g/cm}^3$ at the 30–40 cm depth. The upper layers of the soil (0–30 cm) had an average organic content of approximately 24%, while the bottom layer had an organic content of only 8%. The Upper site is also closest to the headwaters of Strawberry Creek where stagnant water frequently pools on the adjacent cornfield, which would help to slow decomposition, increasing the carbon content and porosity of the Upper soil (Brady and Weil, 1999).

The second study site, referred to herein as the “Lower site”, is located in the southern end of the basin and contains riparian vegetation and the soil at this site is also characterized as a sandy loam, with an average bulk density of approximately $1.06 \pm 0.09 \text{ g/cm}^3$, and a porosity of 53%. The soils at the Lower site were composed of 8% organic material over most of the profile.

3. Methodology

Lysimeters have been used extensively to measure the net movement of water across the soil-atmosphere boundary (Tanner, 1967; Seyfried et al., 2001). A lysimeter is an artificially enclosed volume of soil that can be placed in the field and filled with representative soil and/or vegetation (Brutsaert, 1988). There are two types of lysimeters: (1) a non-weighing lysimeter in which changes in soil water content are measured indirectly; and (2) a weighing lysimeter in which soil water changes are measured by mass, which is the type used in this study.

For this study lysimeters were constructed from a rectangular enclosure (1090 cm^3) nested inside a second plastic container. Material from the surface (including soil and vegetation) was extracted (by block cutting) and placed in the first enclosure. In order

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to maintain the same mechanical properties of the soil, it was placed in the bucket as an undisturbed block (monolith), as suggested by Brutsaert (1988). The enclosure containing the soil/plant monolith had holes drilled into the bottom to facilitate the drainage of water, which was then collected by the second enclosure. Each lysimeter was placed in the hole from which the monolith was extracted.

At the upper and lower sites duplicate lysimeters of bare soil and vegetation were installed (Fig. 1). The first location was at the northern end (Upper site) of the basin located in a riparian area adjacent to a corn field, where one lysimeter contained riparian vegetation and the other bare soil. The second location was also in a riparian area (adjacent to a soybean field) in the lower portion (Lower site) of the basin.

Each lysimeter was weighed with a precision Chatillon® spring balance (Ametak Inc.). The weight of each lysimeter enclosure was recorded and the water that had drained (if any) into the bottom enclosure was also measured by emptying the contents into a one litre graduated cylinder. Soil moisture was also recorded in each lysimeter, and in the immediate area surrounding the lysimeter using a portable Time Domain Reflectometry (TDR) instrument (Hydrosense, Campbell Scientific, Ltd.). This allowed the soil moisture to be monitored to ensure that the soil in the lysimeters was comparable to the surrounding soil. AET from lysimeters is quantified via,

$$AET = P - (V_L + V_R + \Delta V_S) / A, \quad (1)$$

where P is precipitation (mm), V_L is the volume of drainage loss (mm³), V_R is the volume of net surface water exchange (mm³), ΔV_S is the change in the water storage in the lysimeters (mm³), and A is the area of the lysimeter (mm²) (Tanner, 1967). ΔV_S was measured by collecting the water that drained through the contents of the lysimeter and differences in weight of the lysimeter. Lysimeter AET obtained in this manner is then compared with modelled PET values for the same sites.

Automated meteorological stations were installed at the Upper and Lower sites to continuously monitor net radiation, air temperature and relative humidity, wind speed and direction, ground heat flux, and ground temperatures. Only the Lower station was

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equipped with a tipping bucket rain gauge, which was used to record basin precipitation to be used in the lysimeter calculations. Soil moisture TDR probes were also buried 10 cm and 50 cm below the surface at both sites to assess moisture conditions.

5 A vegetation count and survey was conducted in the immediate vicinity of each lysimeter and meteorological station. The area of the vegetation survey was determined by the height of the net radiometer, and the ratio of 10:1 m, where a circular area (or tower footprint) with a radius of 10 m is observed by a radiometer at a height of 1 m (Oke, 1987). The height of the radiometers were 1.5 m and 2 m at the Lower and Upper sites, respectively. Therefore, a radius of 15 m and 20 m was used at the Lower and Upper sites, respectively. The vegetation area within these footprints was recorded
10 along with the coverage of bare soil and then used to areally weight the lysimeter values.

The Priestley-Taylor equation is classified as a radiation-based approach to estimating ET, using net radiation and air temperature, to evaluate equilibrium evaporation, which assumes that an air mass moving over a homogeneous, well-watered surface
15 would become saturated (Dingman, 1994; Priestley and Taylor, 1972). Under these ideal conditions ET would eventually reach a rate of equilibrium (equilibrium potential evapotranspiration, PET_{eq}) (Priestley and Taylor, 1972). Radiation is a very effective parameter to use in measuring equilibrium evaporation or PET. In a review of 30 studies
20 it was commonly found that, in vegetated areas with very small, or no, water deficits approximately 95% of the annual evaporative demand was supplied by radiation (Stagnitti et al., 1989). The Priestley-Taylor model obtains PET_{eq} via,

$$PET_{eq} = \frac{\Delta}{\Delta - \gamma} (Q^* - Q_G), \quad (2)$$

25 where Δ is the slope of the saturated vapour pressure curve ($^{\circ}\text{C/kPa}$), γ is the psychrometric constant (Pa K^{-1}), Q^* is net radiation (Wm^{-2}), Q_G is the soil heat transfer (Wm^{-2}). PET_{eq} is related to AET via the Priestley-Taylor coefficient (α). The term α is

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generally solved using,

$$\alpha = \text{AET}/\text{PET}_{\text{eq}}, \tag{3}$$

where AET is the total measured evaporation and PET_{eq} is the total equilibrium evaporation (Wilson and Baldocchi, 2000). However, equilibrium rarely occurs, as there is almost always horizontal advection and deviations from a “wet” surface (Wilson and Baldocchi, 2000).

In many studies 1.26 is the value used for α , which is based on conditions of minimum advection and no edge effects (Dingman, 1994; Jacobs et al., 2002). However, general estimates for α have been made with respect to certain surfaces. For example, Morton (1983) suggested that an α value of 1.32 should be used for vegetated surfaces due to surface roughness, and in arid regions α should be increased to 1.74. However, a daily average value of 1.26 has been found to be appropriate for most humid climates (Priestley and Taylor, 1972; Rouse and Stewart, 1972).

4. Results

4.1. Evapotranspiration

Precipitation and energy balance components varied very little within the basin, yet evapotranspiration differed between the two sites. The total PET for the Lower site was 201.1 ± 0.1 mm, and for the Upper site was 204.9 ± 0.1 mm. However, the AET at the two sites, obtained using the lysimeters, differed significantly. The Lower site had a total AET of 82.9 ± 3.9 mm, whereas the Upper site had a total of 111.5 ± 5.4 mm. At both sites AET from the vegetation contributed the bulk of the AET. These seasonal AET rates translated to average daily AET rates of 1.05 and 1.14 mm d^{-1} for the Lower and Upper sites, respectively (Fig. 2).

Using the lysimeter data along with energy balance measurements, the Priestley-Taylor coefficient (α) was determined for both sites. As observed with the AET values,

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the α values also varied between the two sites. The α values used are a compilation of all the vegetation in the area and the soil, weighted accordingly to their respective land cover within the footprint of the respective net radiometers. Therefore, over the study period α averaged 0.72 at the Lower site, and 0.88 at the Upper site. A wide range of alpha values were observed over the study period (Fig. 3). At the Lower site the values ranged from 0.26 to 2.04 and at the Upper site the values ranged from 0.08 to 2.03.

4.2. Basin hydrology and climatology

The temperature of the basin is presented as an arithmetic mean of temperature data from three meteorological stations placed throughout the basin (Fig. 1). The mean temperature of the basin over the study period was 18.9 ± 2.6 °C, which is comparable to the 30-year normal of 18.6°C over the same three-month period (Environment Canada, 2004). The temperature over the study period ranged from a low of 12.0°C on 4 June, to a high of almost 24.0°C on 4 July, after which temperatures generally decreased over the course of the season (Fig. 4a). However, during this period of decline a second maximum of 22.8°C was recorded for a brief period on 13 August.

As observed with temperature, the two sites followed a similar seasonal pattern for relative humidity (RH) (Fig. 4e). The Lower site had a slightly higher mean RH of $82.4 \pm 19.3\%$, whereas the Upper site had a mean of approximately $77.0 \pm 17.9\%$ (Fig. 4e). Daily averages for the Lower site ranged from a minimum of 41.8% to a maximum of 100%. For the Upper site, the daily values ranged from 51.9% to 94.9%.

A total of 178 mm of precipitation was recorded over the study period, with the largest single day total of 16.5 mm occurring on 15 July (Fig. 4f). This is considerably lower than the 30-year normal of 259.4 mm over the same time period (Environment Canada, 2004). Precipitation in the basin was classified into three wet periods separated by distinct drier periods noted by a lack of rainfall (Fig. 4f). The three wet periods spanned 2–18 June, 5–21 July and 2–16 August, respectively.

Increases in discharge for the basin followed the large precipitation events at the start and end of the study period. In general, throughout much of the middle of the

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study period the discharge was low and responses to periodic precipitation events were minimal. However, discharge from the middle wet period did not respond in the same manner as the other two rain events (Fig. 4d). This lack of a response in discharge can be explained by dry soil moisture conditions in the basin during that period, where much of the precipitation went into storage or AET. Soil moisture data indicate that both sites had a significant potential for moisture to be stored in the soil during the middle of the season (Figs. 4b and 4c). Figures 4b and 4c show that the upper soil layers were much drier than the deeper layers during this middle period. Thus, during this time the soils had dried enough that the middle wet period produced no response in soil moisture in the lower layer, only the upper layer was affected as soil moisture storage was replenished. During this middle wet period the rain infiltrated rapidly to raise the water table, rather than entering the stream as discharge.

Towards the end of the study period the soil moisture in the upper layers of the soil increased at both the Upper and Lower sites (Figs. 4b and 4c). Over a four-day period (2–6 August) the soil moisture increased from 0.22 to 0.38 and 0.21 to 0.35 at the Lower and Upper sites, respectively. This was followed by a decrease during a reprieve in the precipitation, and then an increase to the highest moisture level of the study period in response to the second largest rain event of the season (12 August). The Lower site reached maximum values of 0.40 and 0.43 for the deeper and shallower layer, respectively, while the Upper site reached 0.39 and 0.44 for the deeper and shallower layer, respectively (Figs. 4b and 4c).

4.3. Energy balance

Figures 5a and 5b show similar net radiation (Q^*) and ground heat flux (Q_G) values for the two sites. The mean Q_G was 0.09 and 0.10 W m^{-2} at the Lower and Upper sites, respectively, and ranged from approximately -0.2 to 0.55 W m^{-2} at both sites over the study period. Mean Q^* values were 106.3 and 100.2 W m^{-2} at the Lower and Upper sites, respectively. Both the Q^* and Q_G values at the two sites showed little

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variability over the study season, but exhibited strong daily fluctuations. The lack of a clear seasonal trend is the result of the study period not spanning the shoulder seasons (spring and autumn).

Wind speed varied the most between the two sites, and was much higher at the Upper site than at the Lower site (Fig. 5d). The average wind speed at the Lower site was $1.5 \pm 1.1 \text{ m s}^{-1}$, and the Upper site was $1.7 \pm 1.4 \text{ m s}^{-1}$ (a difference of 13%). However, the maximum daily wind speed was 3.2 m s^{-1} and 5.7 m s^{-1} for the Lower and Upper sites, respectively. Furthermore, the Upper site wind gusts were observed to exceed 9 m s^{-1} , while at the Lower site the highest wind gust observed was 6.6 m s^{-1} .

Wind direction varied little between the two sites (Fig. 6), with the predominant wind direction being from the southwest. Other wind directions that were less dominant, but occurred equally, were observed from the southeast, south and west. No observations were recorded, at either site, for winds from the north or northeast for any significant amount of time.

5. Discussion

AET rates and α values obtained in this study are comparable with other studies on similar agricultural grass and riparian areas. The daily AET rates of 1.05 and 1.14 mm/day for both sites are well within the range of 0.5 to 4 mm/day reported in a similar terrain by Unland et al. (1998). Measurements of α are a useful tool for irrigation planning purposes, and as such many studies have presented values for a variety of agricultural land-use types. While few have focused primarily on riparian areas the range reported here (0.08 to 2.03) is also well within the range of 0.06 to 2.10 presented for various agricultural grass-dominated sites (Ritchie, 1972; Wallace and Holwill, 1997; Katal and Parlange, 1992; Hares and Novak, 1992b).

The α values also differed between the two sites. The controlling factor of α is moisture availability (Dingman, 1994). However, the precipitation is similar between the Lower and Upper sites, and as such soil moisture and storage explain the differences

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in α .

The data show that the amount of moisture that could be stored in the soil was highest in the middle of the season. Immediately prior to this period, there was a lack of rain allowing the soil to dry out. During the next wet period, the precipitation was able to infiltrate the soil fairly rapidly and remain there, rather than being discharged throughout the basin as lateral groundwater flow (Fig. 2d). During this second wet period the soil moisture graphs indicate that there was more potential for moisture storage at the Upper site than at the Lower site, which can be attributed to the soil properties of the site (high porosity and a high carbon content). This also suggests that more water was available at the Upper site for AET. The α values for the Upper site support this, as they are higher than at the Lower site, indicating that the AET at the Upper site was not as moisture-limited, which seems to be the case at the Lower site. That is, the soil properties at this site permitted easier access to soil moisture for atmospheric exchange.

Although the energy balance at the Lower and Upper sites are very similar, the rates of AET are much higher at the Upper than the Lower site (38% greater). As the other controlling meteorological variables are also similar between the two sites, the only variable that significantly differs between the Lower and Upper sites are the wind parameters (speed and direction), which directly affects turbulent transport, possibly accounting for the observed differences in AET.

Wind speed has already been shown to be a determining factor in estimating evaporation (Thomas, 2000). In a study undertaken in China, Thomas (2000) illustrated that PET was controlled primarily by wind. In desert areas, and in northern areas of the country wind was a more important factor than temperature or relative humidity alone. In desert regions wind is a dominant influence because it is constant force in such regions. In addition, the advection of hot air from the surrounding bare lands could create an "oasis effect" affecting PET rates more than under undisturbed conditions (Thomas, 2000).

The intensity of potential turbulent transport was much higher at the Upper site than

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the Lower site, where wind gusts exceeded 9 m s^{-1} and 7 m s^{-1} , respectively. Daily averages of wind speed were also higher at the Upper site. However, AET was negligible at both sites when wind speeds were lower than 1 m s^{-1} . With wind speeds below 1 m s^{-1} evaporation from the surface leads to a build-up of the moisture content of the air immediately adjacent to the surface (Brutsaert, 1988). When wind speeds increase, drier air is able to replace the saturated air, therefore, horizontal advection is an important process to introduce drier air and increase ET (Brutsaert, 1988). Consequently, with an increase in advective activity, the α will increase.

Wind speed was higher at the Upper site because of the surrounding vegetation. At this site the fields to the southwest, west, northwest and north consisted of short grasses, which was routinely cut, leaving a low and uniform ground cover. Immediately following a cut, bare soil was exposed, and the remaining vegetation was sparse. Under these conditions warmer and drier air would be advected over the location of the Upper lysimeters increasing AET values here relative to the Lower site.

At the Lower site the areas to the north and northwest of the station contained corn and dense riparian grasses, which due to increased friction would reduce wind speeds. This explains why AET was highest at the Lower site when the wind direction was from the south, southwest and west (however, these directions only account for 69% of the time).

6. Conclusions

AET is quite variable within a single terrain unit within a small catchment. This data illustrates that AET is greater at the site in the upper portion of the basin. Therefore, even within the short distances of a small agricultural watershed spatial variability between the two riparian zone sites exists. This further emphasizes the difficulty in quantifying AET rates in multiple land use basins, such as the Strawberry Creek watershed, although it is composed primarily of agricultural crops and similar riparian vegetation along the stream reach. In the Strawberry Creek watershed the primary reasons for

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the difference in AET rates between the sites are differences in wind speed and direction, which also interacts with the differing vegetation of the area to produce contrasting turbulent regimes.

The importance of advective influences and the location's turbulence regime is also illustrated through the differing α values within this basin. The α values at the two sites also indicate that AET at the Lower site is more moisture-limited than the Upper site. This is reflected in the soil parameters at each site, which affect moisture storage that is higher at the Upper site.

Thus, in order to accurately quantify riparian zone AET, it is important to first examine the study site on smaller scales as it has been demonstrated that ET rates can be highly variable at these scales, and must be evaluated as such. While rigorous study of larger sites may be difficult, less intensive research methods could be used such as the Priestley-Taylor model, which can be carried out using automated instruments, once relationships between PET and AET are quantified for the various representative land-use types comprising the system. However, in any study that requires quantifying AET values the controlling parameters of spatial variability must be taken into account, even within the seemingly homogeneous land-use types in a catchment.

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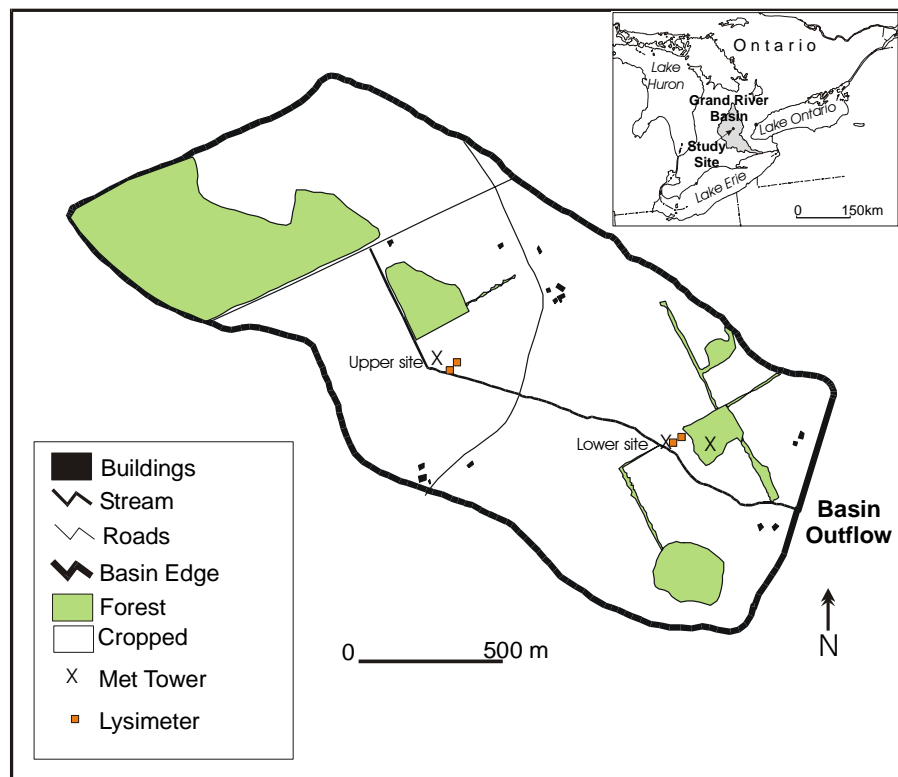


Fig. 1. Schematic of the Strawberry Creek basin, showing the location of instrumentation and land-use types. Inset: Location of the Strawberry Creek study site, relative to Southern Ontario.

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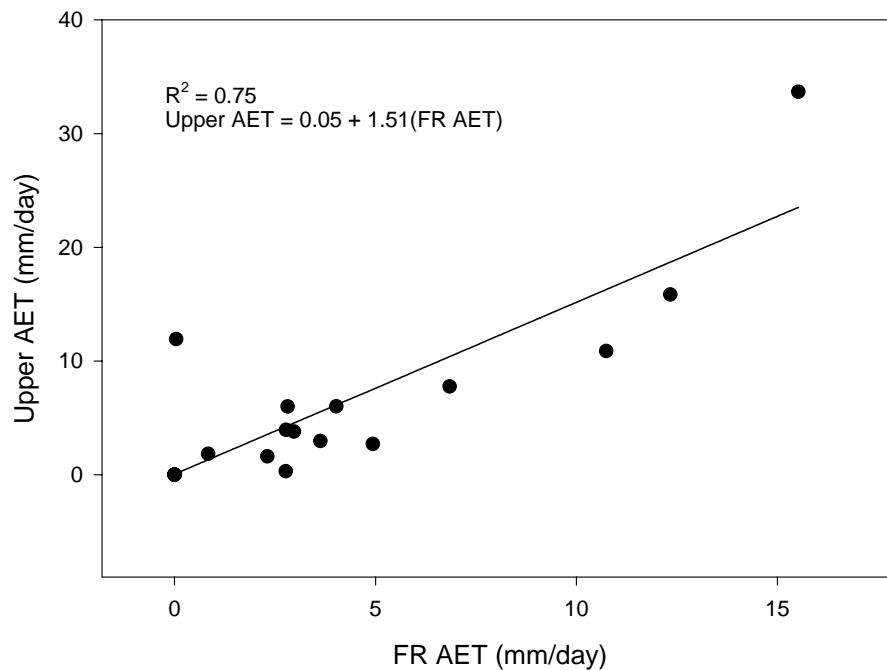


Fig. 2. Actual evapotranspiration (AET) at the Upper and Lower sites, 2 June 2 to 22 August 2003. Strawberry Creek, Maryhill, Ontario.

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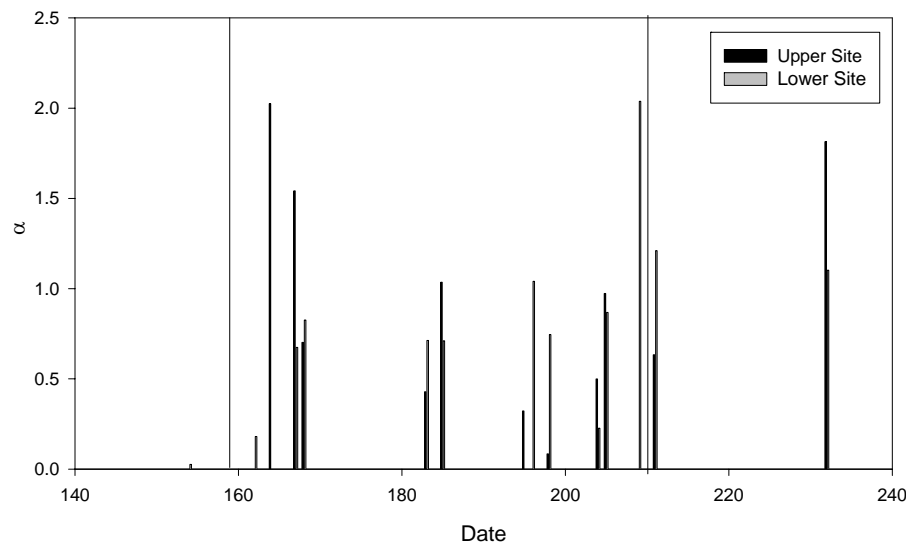


Fig. 3. Priestley-Taylor Coefficient (α) values at the Lower and Upper sites, 2 June to 22 August 2003. Strawberry Creek, Maryhill, Ontario. Vertical lines delineate the wet periods.

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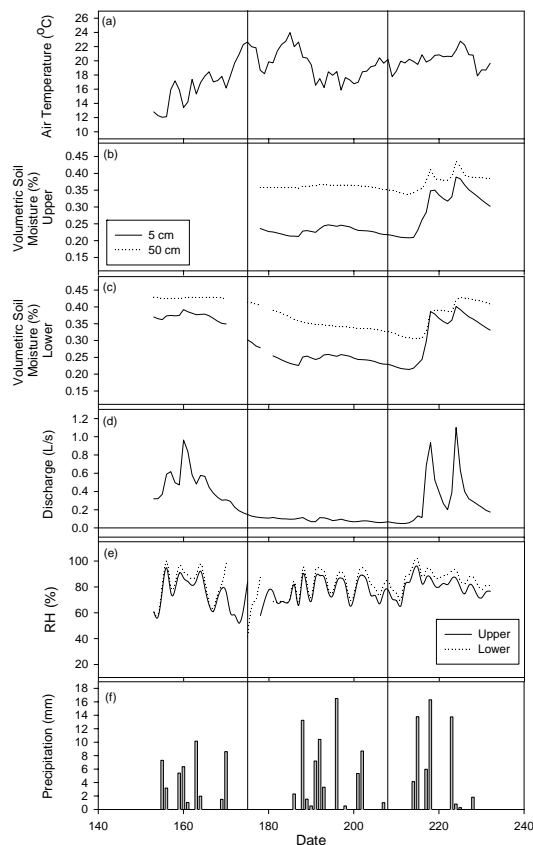


Fig. 4. Basin hydrology and climatology for Strawberry Creek, 3 June to 22 August 2003. Maryhill, Ontario. **(a)** Mean air temperature for the basin, **(b)** Volumetric soil moisture (%) at the Upper site, **(c)** volumetric soil moisture (%) at the Lower site, **(d)** Discharge of the basin (mm), and **(e)** Precipitation (mm). Vertical lines delineate the three wet periods.

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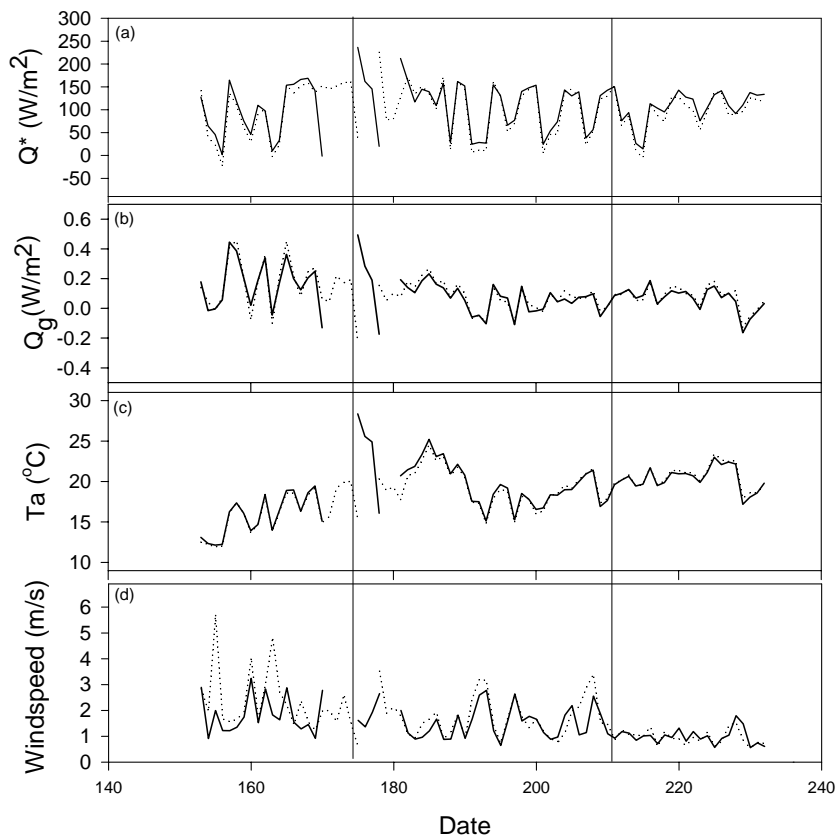


Fig. 5. Energy balance and micrometeorological components for the study period 3 June to 22 August 2003. Strawberry Creek, Maryhill, Ontario. **(a)** Net radiation (W/m^2), **(b)** Ground heat flux (W/m^2), **(c)** Temperature ($^{\circ}\text{C}$), **(d)** Relative Humidity (%), and **(e)** Wind speed (m/s). Data from the Lower site are represented by a solid line and the Upper site by a dotted line. The vertical lines delineate the wet periods.

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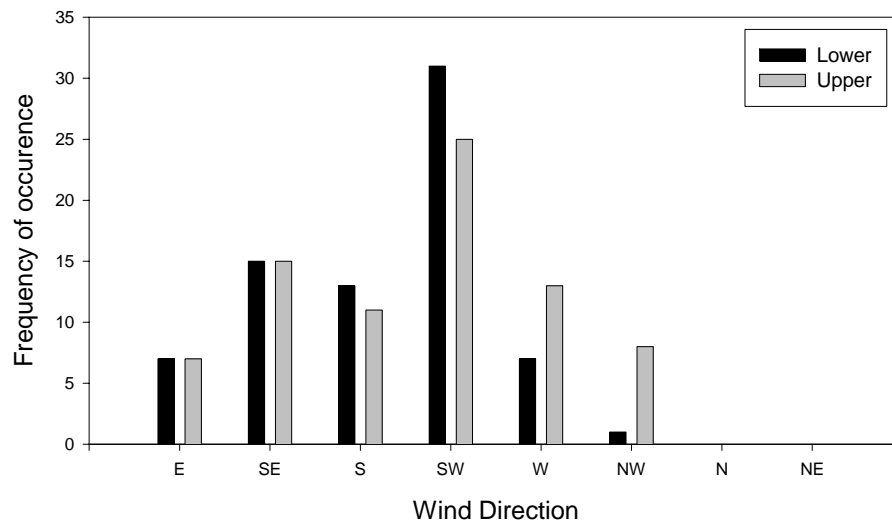


Fig. 6. Wind direction histograms for the Upper and Lower sites, 2 June to 22 August 2003. Strawberry Creek, Maryhill, Ontario.

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